

Ectomycorrhizal fungal communities of silver-fir seedlings regenerating in fir stands and larch forecrops

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Abstract

Key message The diversity of ECM communities of 1-year-old silver-fir seedlings regenerating in mature silver-fir stands is significantly higher than in neighboring larch forecrops.

Abstract Forecrop stands provide the necessary shade for shade preferring seedlings, such as silver-fir, which cannot be introduced as the first generation in open areas. Larch is a good candidate, recommended to be utilized as forecrop. Since fungal symbionts of *Abies alba* seedlings regenerating under larch canopy have not been investigated, we aimed to evaluate the diversity of ECM of 1-year-old silver-fir seedlings regenerating under canopy of larch and to compare these communities to those found in adjacent mature silver-fir stands. Three silver-fir stands (*F*) and three European larch forecrops (*L*) were selected to study. Seedling size did not differ between *F* and *L* stands. ECM colonization was observed in 100% silver-fir seedlings in *F* and 80% in *L* stands. The fine roots were highly colonized, and 91 and 87% of live ECM were found in seedlings from *F* and *L* stands, respectively. Sequencing analysis revealed a total 53 ECM taxa. The observed number of ECM taxa in

the *F* stands was significantly higher (46) than that in the *L* forecrops (25), and 34% of taxa were common to both stands. The dominant ECM species in *F* were unidentified fungus 1, *Piloderma* sp., *Tylospora asterophora* and *Russula integra*. Fir seedlings regenerating in *L* forecrops formed ectomycorrhizas mostly with unidentified fungus 1, *Tomentella subtilacina*, *Tylospora* sp., *Hydnotrya bailii* and *T. asterophora*. Based on ANOSIM analysis, ECM communities have shown significant differences between study sites. The diversity of ECM fungal partners and the high colonization rate of silver-fir seedlings regenerating in larch forecrop stands should be sufficient to provide efficient afforestation of post-arable lands and gives the opportunity for their successful rebuilding.

Keywords Ectomycorrhiza · *Abies alba* seedlings · Fir stands · Larch forecrops · ITS rDNA

Introduction

Since the middle of the last century, some of the former arable lands in the Carpathians (Poland) have been afforested with forecrop stands. These stands provide the necessary shade for low light (shade preferring) seedlings, such as *Abies alba*, which cannot be introduced as the first generation in open areas. Silver-fir is one of the most important forest trees in the mountainous regions of Central Europe (Jaworski 2011). One of the essential environmental factors that determines seed germination (Bormann 1983) and plant growth and subsequently regeneration and survival of silver-fir is the understory light regime. Seedlings and saplings require only 15–25% of full light (Jaworski 2011), which determine the methods available for regeneration and cultivation during the juvenile stage of the fir beneath

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Table 1 Localization, tree species composition, and soil parameters of the examined silver-fir and larch stands

	Silver-fir stands			Larch forecrops		
	F1	F2 ^a	F3 ^a	L1	L2	L3
GPS	49°26'27.9"N 20°57'18.2"E	49°26'23.6"N 20°57'37.6"E	49°21'26.7"N 20°59'30.8"E	49°26'26.7"N 20°57'14.2"E	49°26'22.1"N 20°57'40.1"E	49°21'25.9"N 20°59'41.0"E
Tree species composition	Silver-fir (100%)	Silver-fir (100%)	Silver-fir (100%)	Larch (70%), Pine (20%), Spruce (10%)	Larch (90%), Spruce (10%)	Larch (100%)
Tree age (years)	45–140	45–130	75–95	50	55	50
Area (ha)	19.9	6.3	5.4	1.7	3.2	8.1
Soil type	Acid brown (cambic)					
pH in H ₂ O	4.1	4.3	4.3	4.0	4.0	4.5
pH in KCl	3.3	3.4	3.3	3.1	3.2	3.5
C (%)	3.78	3.58	4.38	4.15	3.75	3.38
N (%)	0.24	0.24	0.28	0.25	0.22	0.24
C/N	15.8	15.0	15.5	16.3	16.8	14.3
Ca (mg/kg)	165.6	302.7	441.3	187.3	197.4	468.0
K (mg/kg)	64.2	57.8	80.6	53.7	50.2	93.2
Mg (mg/kg)	26.6	39.7	59.2	34.9	37.3	74.3
Na (mg/kg)	7.5	8.7	6.8	8.1	7.1	6.9
P (mg/kg)	1.3	1.3	1.2	1.9	1.6	2.6

^aData published by Ważny (2011, 2014)

the understory. Among tree species recommended to be utilized as a forecrop, larch is one of the best known candidates due to its ability to provide appropriate light conditions for fir seedlings. It was documented that the optimal conditions for silver-fir seedlings development in the Karkonosze Mountains (Poland) were found under larch canopy (Dobrowolska 2008), demonstrated by the highest increment of height, diameter, and offshoots. Photosynthetically active radiation (PAR) reached 20% under larch crowns, which positively affected growth of fir needles (Robakowski et al. 2004). Spruce (*Picea abies*) and beech (*Fagus sylvatica*) forecrops are unsuitable, because they do not provide recommendable light conditions for fir regeneration (Dobrowolska 2008). Taking into account this aspect larch has become one of the most popular forecrop tree species in the middle of the 20 century for the afforestation of the post-agricultural areas in the Polish Carpathians. Currently, these stands have reached the rebuilding phase—mature trees provide habitat for shade preferring silver-fir and these forests are gradually being converted by natural seed regeneration of silver-fir from adjacent stands or planting fir seedlings under the larch canopy. Rebuilding of forecrop stands may take 30–50 years.

Abies alba is an obligate mutualist with ectomycorrhizal (ECM) fungi (Trappe 1962). Formation of ECM on tree roots protects them against root diseases, ensuring healthy seedling growth (Marx 1969) and increases the survival rate of silver-fir seedlings (Kowalski 1982). Studies of ECM communities of silver-fir grown under Scots pine (*Pinus sylvestris*) canopy

documented that the diversity of ECM symbionts was lower in these stands than in mature silver-fir stands (Farfał 2008; Kowalski et al. 1996; Ważny 2014). To our knowledge, the fungal symbionts of *A. alba* seedlings regenerating under larch canopy have not been investigated. Moreover, the mycobionts of *A. alba* ectomycorrhizas are still poorly identified. The current knowledge of silver-fir symbionts is predominantly based on morphological and anatomical descriptions of ectomycorrhizas (Agerer 1987–2007; Berndt et al. 1990; Comandini et al. 2001; De Román et al. 2005; Dominik 1961; Farfał 2008; Kowalski 1982; Pachlewski 1955; Stepniewska and Rębisz 2004) and fungal fruiting bodies (Laganà et al. 1999, 2002). Only a few molecular investigations of the ECM symbionts that associate with silver-fir have appeared in the literature (Cremer et al. 2009; Eberhardt et al. 2000; Rudawska et al. 2015; Ważny 2014). In our study, we aimed to evaluate the diversity of ECM communities of 1-year-old silver-fir seedlings regenerating under canopy of larch and to compare these communities to those found in adjacent mature silver-fir stands.

Materials and methods

Study sites

Six stands located in the Experimental Forestry Unit in Krynica (Poland) were selected for the study: three silver-fir (*A. alba*) stands (F1, F2, and F3) and three larch

(*L. decidua*) stands (*L1*, *L2*, and *L3*) with natural regeneration of silver-fir seedlings. Their detailed description is presented in Table 1. Larch stands were treated as forecrop stands. The paired, *L* and *F*, stands were neighbored. Each larch forecrop was established on an abandoned area after cultivation. The results from site *F2* and *F3* were presented in the previous papers by Ważny (2011, 2014) and are included here due to the fact that fir stands were used as a reference fir stand for *L2* and *L3* forecrop stands. For ECM analysis, 30 1-year-old seedlings were sampled along parallel transects spaced 10–20 m apart with minimum intervals of 10 m between particular seedlings. To avoid edge effect in fir stands, seedlings were collected at least 20 m from the stand edges. If the seedlings were not abundant in forecrop stands, minimum intervals (10 m) between samples were omitted. All the seedlings sampled in fir stands were the result of natural regeneration. In the case of larch stands, the number of fir seedlings naturally regenerated was not always enough and sampled seedlings were also the result of artificial regeneration by direct seeding. In forecrops, the places with other tree species were excluded from sample collection. The seedlings were collected together with adjacent soil surrounding the roots (approx. 250 ml), placed in plastics bags, and stored at -20°C until analysis. In each stand, one bulk soil sample (each being composed of ten subsamples) was collected for chemical analysis.

Seedling parameters

Average values of a number of mycorrhizal seedlings, number of alive and dead ectomycorrhizas and non-mycorrhizal roots, height and diameter of shoot, and dry weight of shoot and root were assessed.

ECM assessment

The root system was gently washed in tap water to remove organic and mineral matter. All the root tips per seedling were assessed for mycorrhizal colonization (2520 root tips per 180 seedlings). Ectomycorrhizas were described according to Agerer (1987–2007) based on morphology features (color, shape, and texture of mantle, ectomycorrhiza ramification, presence and appearance of extramatrix hyphae, rhizomorphs, and cistidia). For each morphotype from each site, two-to-six root tips were transferred for molecular identification of the fungal symbiont. DNA extraction was performed according to the modified method by Lanfranco et al. (1998). Amplification of the internal transcribed spacer (ITS) rDNA region was carried out with ITS1F (Gardes and Bruns 1993) and ITS4 (White et al. 1990) primers. The polymerase chain reaction (PCR) was performed as follows: $1\times$ Pol Buffer; 0.2 mM dNTPs (each); 50 pmol of each primer; and 1 unit Taq DNA polymerase (Eurx). Amplification was carried out as follows: initial denaturation at 93°C for 10 min followed by 35 cycles of 1-min denaturation at 95°C , 1-min annealing at 60°C , 2-min extension at 72°C , and 10-min final extension at 72°C . PCR products were separated by electrophoresis in 1.5% agarose gel stained with Midori Green. Sequencing was carried out at the Laboratory of Molecular Biology of Adam Mickiewicz University in Poznan, Poland. The ITS4 primer was used for reading sequences. The sequences were edited using BioEdit (Hall 1999) and Chromas (<http://www.technelysium.com.au>) software and then compared with published sequences in UNITE (Abarenkov et al. 2010) and NCBI (<http://www.ncbi.nlm.nih.gov>) databases. A positive identification of a mycorrhizal species was confirmed if they shared $\geq 98\%$ ITS region sequence identity with the most similar (reference) sequence from UNITE or NCBI databases. The obtained sequences within 2%

Table 2 Biometric parameters and mycorrhizal colonization of *Abies alba* seedlings regenerating in fir stands (*F*) and larch forecrops (*L*)

Parameters	<i>F1</i>	<i>F2</i> ^a	<i>F3</i> ^a	<i>L1</i>	<i>L2</i>	<i>L3</i>	<i>F</i> ^b	<i>L</i> ^b
Height (cm)	4.8 ± 0.1	4.8 ± 0.1	5.1 ± 0.1	4.8 ± 0.1	4.9 ± 0.1	4.2 ± 0.1	$4.9 \pm 0.1a$	$4.6 \pm 0.2a$
Diameter (mm)	0.9 ± 0.0	1.0 ± 0.0	1.0 ± 0.0	0.9 ± 0.0	0.9 ± 0.0	1.0 ± 0.0	$1.0 \pm 0.0a$	$0.9 \pm 0.0a$
Dry weight (g)	0.028 ± 0.003	0.034 ± 0.003	0.040 ± 0.003	0.025 ± 0.003	0.026 ± 0.003	0.030 ± 0.003	$0.033 \pm 0.003a$	$0.027 \pm 0.001a$
Mycorrhizal seedlings (%)	100.0	100.0	100.0	80.0	70.0	90.0	$100.0 \pm 0.0b$	$80.0 \pm 5.8a$
Live mycorrhizas (%)	87.3	86.4	100.0	96.5	68.5	96.1	$91.2 \pm 4.4a$	$87.0 \pm 9.3a$
Non-mycorrhizal roots (%)	12.7	13.6	0.0	2.8	30.6	3.9	$8.8 \pm 4.4a$	$12.5 \pm 9.1a$
Dead mycorrhizas (%)	0.0	0.0	0.0	0.7	0.9	0.0	$0.0 \pm 0.0a$	$0.5 \pm 0.3a$

^aData published by Ważny (2011, 2014); different letters between *F* and *L* mean statistically significant differences (*t* test, $p \leq 0.05$)

^bData are presented as a mean \pm SE ($n=3$)

Table 3 Relative abundance (RA) and frequency (FR) of ectomycorrhizal taxa on the roots of 1-year-old *Abies alba* seedlings regenerating in fir stands (F1, F2, and F3) and larch forecrops (L1, L2, and L3)

Fungal species	F1		F2 ^a		F3 ^a		L1		L2		L3	
	RA	FR	RA	FR	RA	FR	RA	FR	RA	FR	RA	FR
<i>Amanita muscaria</i>					1.6	13.3						
<i>Amanita rubescens</i>	3.3	3.3										
<i>Boletus edulis</i>					0.7	6.7						
<i>Boletus pruinatus</i>	0.5	6.7	2.0	6.7	3.9	13.3	1.3	6.7				
<i>Cenococcum geophilum</i>	1.9	10.0	5.6	13.3	2.7	16.7					0.3	3.3
<i>Clavulina cristata</i>	11.8	23.3										
<i>Clavulina</i> sp. 1			5.3	23.3	1.4	10.0					12.0	6.7
<i>Cortinarius</i> sp. 1					0.5	10						
<i>Cortinarius</i> sp. 2	5.8	6.7										
<i>Cortinarius</i> sp. 3					1.4	10						
<i>Elaphomyces muricatus</i>	1.4	3.3			2.1	6.7			0.7	3.3		
<i>Entoloma</i> sp.	1.4	3.3										
<i>Hydnotrya bailii</i>							23.1	20.0				
<i>Hydnotrya</i> sp.					2.8	6.7						
<i>Hydnotrya tulasnei</i>									2.9	10.0		
<i>Hydnum rufescens</i>					2.3	3.3						
<i>Laccaria laccata</i>											7.7	13.3
<i>Laccaria maritima</i>									1.4	3.3		
<i>Lactarius aurantiacus</i>					8.2	30.0						
<i>Lactarius lignyotus</i>	1.1	3.3										
<i>Lactarius necator</i>									13.4	23.3		
<i>Leotiomyces</i>					6.2	13.3						
<i>Mycena galopus</i>							2.0	3.3				
<i>Paxillus involutus</i>					1.1	3.3						
<i>Piloderma</i> cf. <i>byssinum</i>					0.9	3.3						
<i>Piloderma fallax</i>	0.3	3.3			2.5	13.3						
<i>Piloderma</i> sp. 1			32.2	20.0								
<i>Pseudotomentella</i> sp. 1			5.1	20.0			1.6	6.7				
<i>Pseudotomentella</i> sp. 2	7.7	6.7							2.2	3.3		
<i>Russula amethystina</i>	7.1	6.7			4.8	20.0						
<i>Russula cyanoxantha</i>			1.0	6.7								
<i>Russula fellea</i>	2.5	3.3										
<i>Russula integra</i>	18.1	26.7	11.1	20.0							4.3	6.7
<i>Russula ochroleuca</i>									15.2	20.0		
<i>Russula olivacea</i>			1.0	3.3	5.7	23.3						
<i>Russula puellaris</i>			0.8	3.3			5.2	6.7				
<i>Russula</i> sp. 1	6.9	6.7	0.3	3.3								
<i>Russula vesca</i>	0.8	3.3										
<i>Sebacina</i> sp. 1	0.8	3.3							3.6	3.3		
<i>Sebacina</i> sp. 2					0.7	3.3						
<i>Sebacina</i> sp. 3	12.1	26.7							2.9	6.7		
<i>Thelephora</i> sp.	1.1	6.7									5.4	6.7
<i>Thelephora terrestris</i>			1.3	3.3			2.9	6.7				
<i>Tomentella stuposa</i>					7.3	30.0						
<i>Tomentella subtililacina</i>			2.5	3.3							45.6	56.7
<i>Tomentellopsis</i> sp.	0.5	6.7	0.5	3.3								
<i>Tylospora asterophora</i>	11.8	40	8.1	20.0	6.8	16.7					17.8	30.0
<i>Tylospora fibrillosa</i>			4.6	10.0			5.9	13.3				
<i>Tylospora</i> sp.			3.0	10.0			20.2	26.7	23.9	30.0		

Table 3 (continued)

Fungal species	F1		F2 ^a		F3 ^a		L1		L2		L3	
	RA	FR	RA	FR	RA	FR	RA	FR	RA	FR	RA	FR
<i>Unidentified 1</i>			13.4	13.3	35.1	70.0	37.8	53.3	33.0	33.3	6.9	16.7
<i>Unidentified 2</i>			2.3	6.7								
<i>Unidentified 3</i>	1.9	3.3										
<i>Xerocomus badius</i>	1.1	3.3			1.1	3.3			0.7	3.3		

^aData published by Wazny (2011, 2014)

nucleotide difference were categorized as a single operational unit and assigned an identical name.

Soil assessment

The pH was determined in H₂O and KCl, the C content was analyzed by the Tiurin method, and the N content by the Kjeldahl method and the macronutrient (Ca, K, Mg, Na) content in the soil was determined in 1 M CH₃COONH₄ with the ICP-OES Thermo iCAP 6500 DUO spectrophotometer (Ostrowska et al. 1991).

Data analysis

Statistical analysis of the seedling biometric data (height, diameter and dry weight) was performed with the parametric *t* test and nonparametric Mann–Whitney *U* test using Statistica 10.0 (StatSoft 2011) at the level of significance $\alpha=0.05$. The nonparametric test was used when Shapiro–Wilk normality and Levene’s homogeneity of variance were not found. To assess if a sufficient number of samples were collected, the observed species accumulation curve and jackknife first-degree estimator curve with 100 randomization with sample replacement were plotted in the Estimates 9.1.0 software (Colwell 2006). Communities of ECM fungi were described by species richness (number of identified mycorrhizal taxa), relative abundance (number of mycorrhizas of a given mycorrhizal species per total number of mycorrhizas in each site), and frequency (ratio of number of seedlings with given ECM taxa to total number of seedlings in each stand). The ECM species diversity was evaluated by Shannon–Wiener’s (*H'*) and Simpson’s (*1D*) indicators. The variability of ECM composition among seedlings of different forest stands was visualized using non-metric multidimensional scaling (NMDS). Analysis of similarity ANOSIM was used to determine if relative abundance of ECM symbionts differed between study sites. The above coefficients, ANOSIM, NMDS, and the cluster analysis were calculated in PAST 2.17 software (Hammer et al. 2001) based on standardized, square root transformed data of each taxa. The Bray–Curtis dissimilarity coefficient was chosen for calculation in analyses (Bray and Curtis 1957).

Results

ECM colonization was observed in 100% silver-fir seedlings in *F* and was significantly higher than in *L* stands (80%). The fine roots were highly colonized, 91 and 87% of live ECM were found in seedlings from *F* and *L* stands, respectively (Table 2).

Sequencing analysis revealed a total 53 ECM taxa on silver-fir seedlings (Table 3). The observed number of ECM taxa in the *F* stands was significantly higher than that in the *L* forecrops; 46 taxa were present on seedlings in *F* stands and 25 taxa were present on seedlings in *L* stands (Fig. 1). Based on the jackknife first-degree estimator, the observed number of taxa was 77% of the estimated richness in fir stands and 78% of the estimated number of ECM species in larch forecrops (Fig. 2). Thirty-three ECM morphotypes were described to the species level. Three morphotypes were unidentified to any taxa. Eighteen out of 53 taxa (34%) were common to both stands. The dominant ECM species identified in seedlings regenerating in fir stands were: unidentified fungus 1 (18.9%), followed by *Piloderma* sp. (9.6%), *Tylospora asterophora* (8.6%) and *Russula integra* (8.3%). Fir seedlings regenerating under larch canopy formed ectomycorrhizas mostly with unidentified fungus 1 (24.8%), *Tomentella sublilacina* (17.1%), *Tylospora* sp. (13.7%), *Hydnotrya bailii* (7.6%), and *T. asterophora* (6.7%) (Fig. 1).

Species richness per site was twofold higher in studied *F* stands (18–22) compared to *L* stands (8–11) (Table 3). Similarly, species richness per one seedling was also higher in *F* stands (1.9–3.3 versus 1.4). The Shannon–Wiener and Simpson’s indices for the ECM assemblages were higher in *F* stands (2.28–2.58 and 0.85–0.90) than in *L* stands (1.61–1.81 and 0.73–0.79), respectively (Table 4).

Based on ANOSIM analysis, ECM communities have shown significant differences ($R=0.2771$, $p=0.0001$) between study sites. Non-metric multidimensional scaling plot grouped *L1*, *L2*, and *F3* sites closely (Fig. 3). This observation was confirmed in the similarity dendrogram for the ECM fungi assemblages in study sites (Fig. 4). Here, the most similar (66%) ECM communities were found in *L1* and *L2* stands. Closely related to this clade was ECM communities noted in *F3* stand.

Fig. 1 Mean relative abundance of ECM fungi on 1-year-old *Abies alba* seedlings regenerating in fir stands and larch forecrops

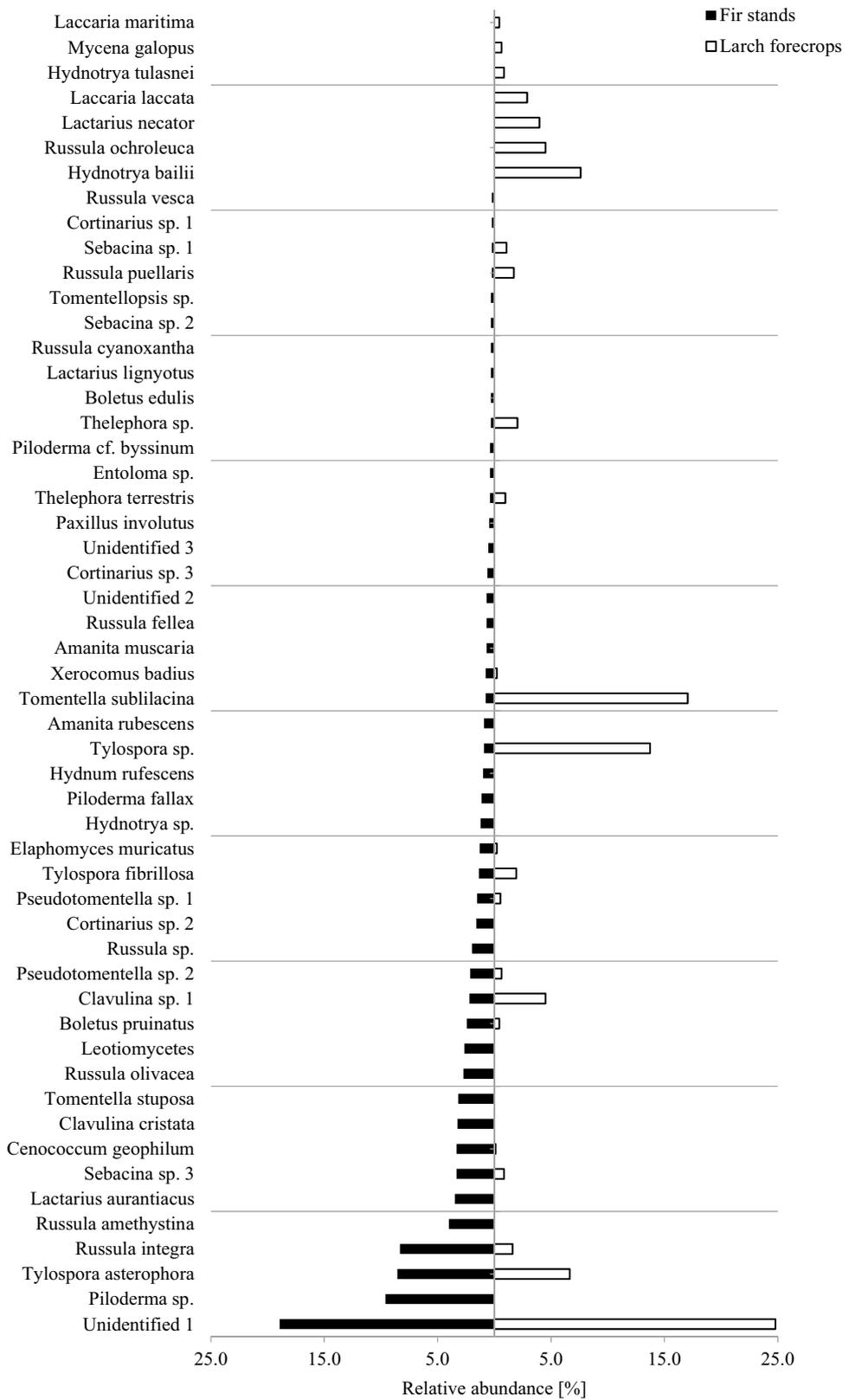


Fig. 2 Ectomycorrhizal species richness estimation curves for 1-year-old *Abies alba* seedlings regenerating in fir (*F*) and larch (*L*) stands. Sobs-species observed, Jack1-first order jack-knife estimator (100 randomized runs with sample replacement were used)

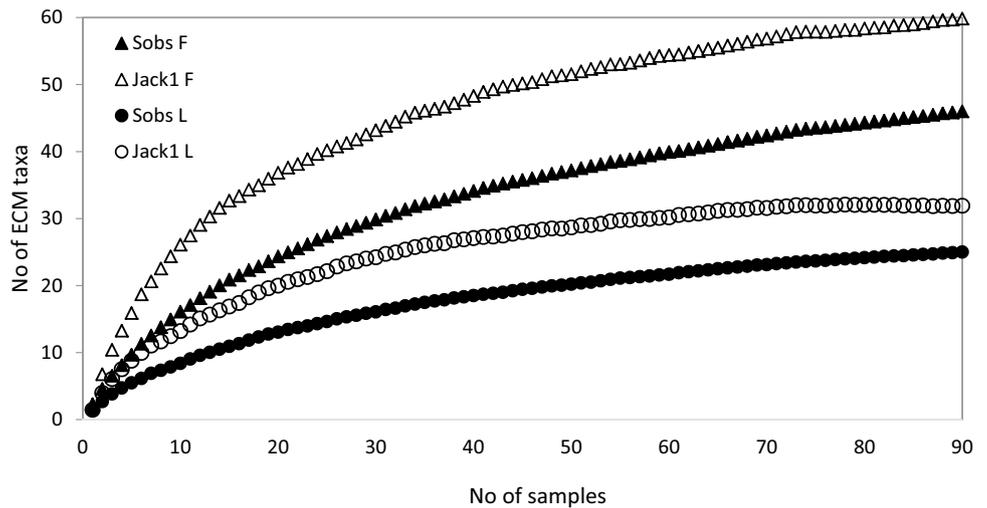


Table 4 Diversity indices of ECM communities colonizing *Abies alba* seedlings regenerating in fir stands (*F*) and larch forecrops (*L*)

Parameters	<i>F</i> 1	<i>F</i> 2 ^a	<i>F</i> 3 ^a	<i>L</i> 1	<i>L</i> 2	<i>L</i> 3	<i>F</i> ^b	<i>L</i> ^b
Richness per site	22	18	22	9	11	8	20.7 ± 1.3b (total 46)	9.3 ± 0.9a (total 25)
Richness per seedlings	2.1	1.9	3.3	1.4	1.4	1.4	2.4 ± 0.4a	1.4 ± 0.0a
Shannon–Wiener (<i>H'</i>)	2.58	2.28	2.44	1.65	1.81	1.61	2.43 ± 0.09b	1.69 ± 0.06a
Simpson (1- <i>D</i>)	0.90	0.85	0.85	0.76	0.79	0.73	0.87 ± 0.02b	0.76 ± 0.02a
Dominance (<i>D</i>)	0.10	0.15	0.15	0.24	0.21	0.27	0.13 ± 0.02b	0.24 ± 0.02a

^aData published by Ważny (2011, 2014); different letters between *F* and *L* mean statistically significant differences (*t* test, *p* ≤ 0.05)

^bData are presented as a mean ± SE (*n* = 3)

Soil parameters are presented in Table 1. The pH value was similar among study sites, ranging from 4.0 to 4.5 in H₂O and from 3.1 to 3.5 in KCl, C/N ratio ranged from 14.3 to 16.8%. Content of K ranged from 1.2 to 1.3 in *F* sites to 1.6–2.6 mg/kg in *L* sites.

Biometric parameters (height, diameter, and dry weight) of the seedlings did not differ between fir (*F*) and larch (*L*) stands (Table 2). However, in the sites with the lowest percentage of mycorrhizal seedlings (*L*1 and *L*2), diameter and dry weight of the seedlings were significantly higher in mycorrhizal seedlings compared to non-mycorrhizal seedlings (data not published).

Discussion

To our knowledge, this is the first report concerning ECM of 1-year-old *A. alba* seedlings regenerating in larch forecrops using molecular approaches. Our study has shown that the diversity of ECM communities of 1-year-old silver-fir seedlings regenerating in larch forecrops is lower than in neighboring mature silver-fir stands. This result supports our previous observation documented in Scots pine forecrops (Ważny 2014). Species richness, Shannon–Wiener

and Simpson’s diversity indices were significantly higher in mature fir stands compared to forecrops. Fungal colonization of fine roots in both stands (*F* and *L*) was very high (91 and 87%, respectively). Species richness in larch forecrops (8–11) was similar to the one reported in Scots pine forecrops (7–12) (Ważny 2014).

Only 34% fungal symbionts of silver-fir seedlings were common to fir and larch stands. The vast majority (66%) of fungal symbiont of silver-fir seedlings in larch forecrops were not common with silver-fir in fir stands and may be shared with coexisting mature European larch. For instance, *Abies homolepis*, closely related to *A. alba*, grown in broadleaf forest initially shared ECM with broadleaf trees and become colonized by specific ECM many years after its establishment (Ishida et al. 2007). Similarly, as it was suggested by Rudawska et al. (2015), tree species grown outside its range are readily colonized by available and compatible ECM fungi. CMN is likely to be the dominant form of infection for regenerating seedlings (Nara 2008).

In this study, we identified 53 ECM taxa [sampling intensity (SI) = 180 samples]. A similar ECM morphotype richness (48) was found on mature silver-fir trees in Apennines (Italy) by Comandini et al. (2001). On

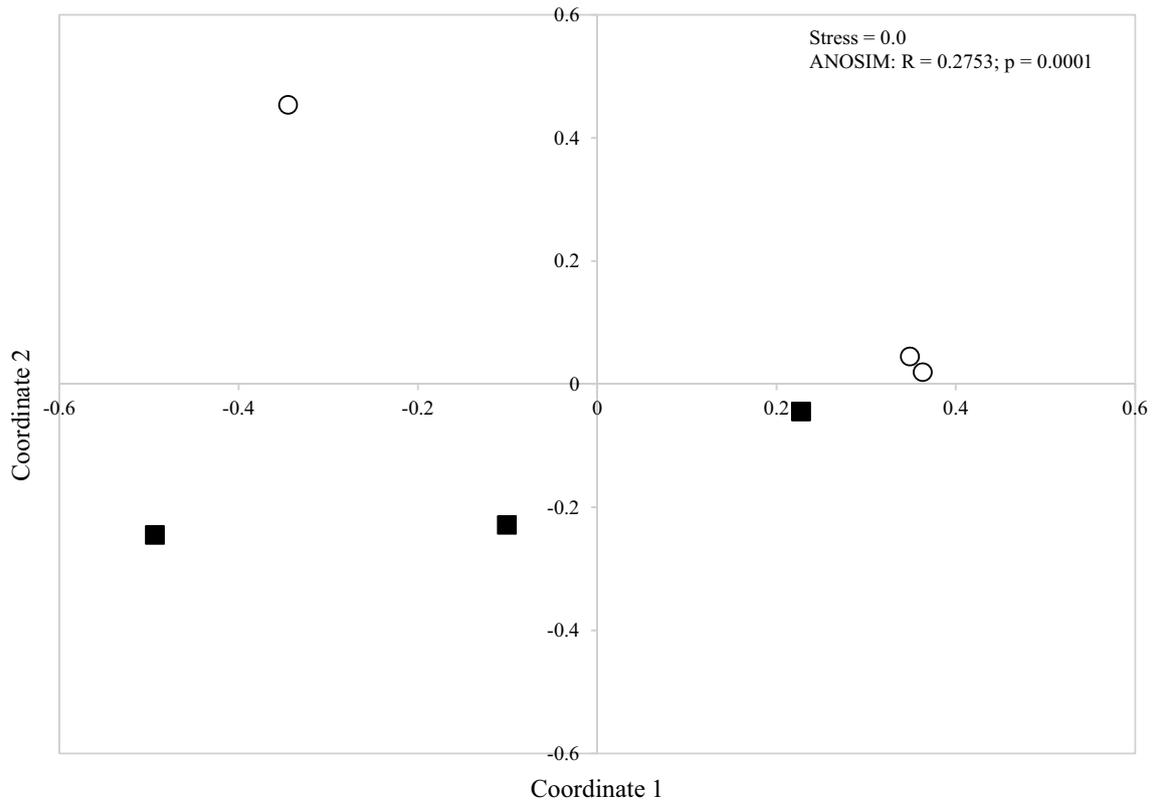
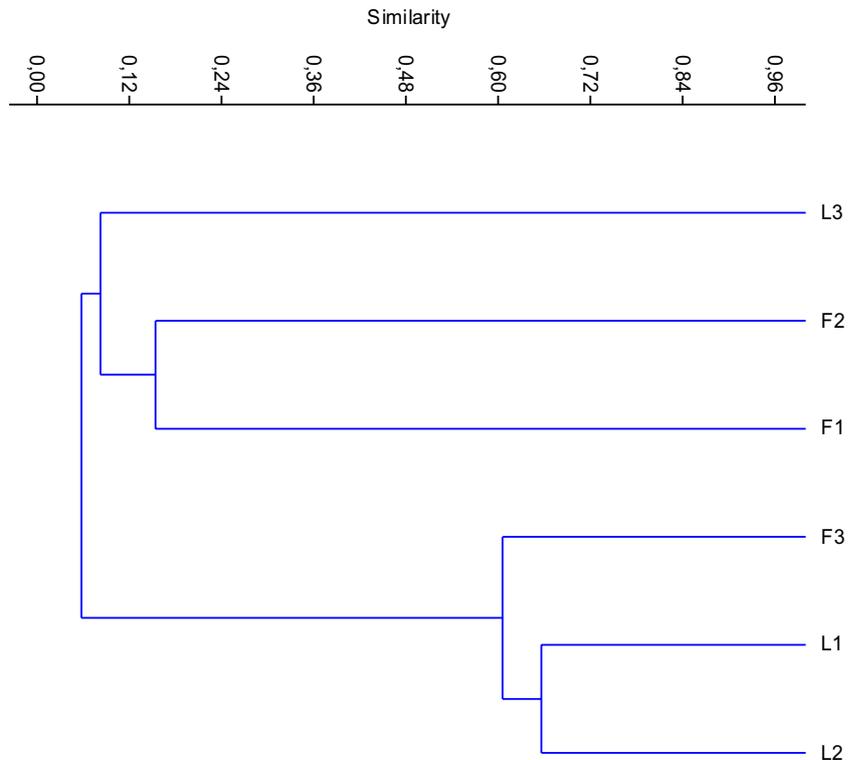


Fig. 3 Non-metric multidimensional scaling plot for ECM fungi communities on 1-year-old *Abies alba* seedlings based on Bray–Curtis distance across sites (*black squares* mean fir stands; *open circles* mean larch forecrops). Each point represents ECM assemblages of 30 seedlings

Fig. 4 Mean similarity dendrogram for the ECM fungi assemblages in study sites



seedlings from the same locality, these authors (Comandini et al. 1998) identified 25 ECM morphotypes (SI=30). Based on anatomical and morphological features, Kowalski (2008) found 35 ECM types on silver-fir seedlings in the Karkonosze National Park (Poland). Based on ECM sporocarp identification, Laganà et al. (1999, 2002) observed over 40 ECM fungal species in silver-fir forests in Tuscany (Italy). However assemblages of ECM fungal symbionts based on aboveground sporocarp identification can be significantly different from belowground ECM communities (data not published). ECM symbionts of *A. alba* are still poorly examined with utilization of molecular approaches. Based on DNA sequencing, sampled directly from ECM root tips, in German forests Cremer et al. (2009) and Schirkonyer et al. (2013) identified 33 and 15 taxa (SI=42) of *A. alba* symbionts, respectively, and Rudawska et al. (2015) observed 35 mycorrhizal fungal taxa (SI=64) on mature *A. alba* outside its natural range in Pomerania (northern Poland). In closely related to silver-fir *Abies lasiocarpa*, Kranabetter et al. (2009) identified 74 ECM symbionts (SI=95). Matsuda and Hijii (1999, 2004) revealed 37 morphotypes from *Abies firma* seedlings (SI=142).

The relative abundance of fir symbionts documented in fir stands in our previous study and presented here was comparable, the relative abundance of fir symbionts documented in pine (Ważny 2014), and larch forecrops was completely different. Fir seedlings from the pine forecrops formed mycorrhizas mostly with *Clavulina cristata*, *Tomentella* sp., *Tuber puberulum* and *Clavulina* sp. However, fir seedlings regenerating under larch canopy formed ectomycorrhizas mostly with unidentified fungus 1, *T. sublilacina*, *Tylospora* sp., *H. bailii*, and *T. asterophora*. Only 6 of 48 fungal symbionts were common to pine and larch forecrops, including: *Clavulina* sp., *Sebacina* sp., *Tomentella stuposa*, *Tomentellopsis* sp., *T. asterophora*, and *Xerocomus badius* (Ważny 2014). Among 25 symbionts of fir seedlings regenerating in larch forecrops, 8 species were documented as symbionts of larch seedlings naturally regenerated under larch canopy documented by Leski and Rudawska (2012): *Boletus pruinatus*, *Cenococcum geophilum*, *Hydnotrya tulasnei*, *Russula ochroleuca*, *R. puellaris*, *Thelephora terrestris*, *T. sublilacina*, and *X. badius*.

The most frequent fungal taxa was unidentified fungus 1 (five sites), followed by *B. pruinatus*, *C. geophilum* and *T. asterophora* which were presented at least in four sites. Unidentified fungus 1 characterized brown mycorrhizas with the Hartig net and without a mantle. The Hartig net can appear before the fungal mantle (Nylund and Unesam 1982). Molecular analysis revealed that this morphotype was formed by various taxa and this is probably

an initially stage of mycorrhiza of many fungal species which differentiates later on. On 2-year-old *A. alba* seedlings, this morphotype was observed; however, was not as frequent (data not published). *Boletus pruinatus* was previously documented as fungal component of ECM on silver-fir (Cremer et al. 2009; Schirkonyer et al. (2013); Ważny 2014) and European larch symbiont (Leski and Rudawska 2012), but it was not revealed as a dominant species. On the other hand, *C. geophilum* is a common, dominating fungal component of the ectomycorrhizas of many tree species (Aučina et al. 2011; Teste et al. 2009), silver-fir including (Cremer et al. 2009; Ważny 2014). Even though the investigation was conducted in *A. alba* natural range, we did not observe *L. salmonicolor*, known to be specific to *A. alba*, which has been already documented in the Carpathians (Ważny 2014), the Apennines (Comandini et al. 1998, 2001), and Tuscany (Laganà et al. 2002).

This is the first report concerning ECM of *A. alba* seedlings regenerating in larch forecrops using molecular approaches. These results shed a new light on the diversity of ECM fungal species associated with silver-fir. The diversity of ECM communities of 1-year-old silver-fir seedlings regenerating in mature silver-fir stands was significantly higher than in neighboring larch forecrops. However, we still do not know, why the diversity ECM of silver-fir seedlings regenerating under larch canopy was different than in mature fir stands. The comparison ECM communities on mature fir and larch trees in the examined forests would provide the necessary insight to solve this problem.

Author contribution statement RW: laboratory works, data analysis, preparing the manuscript; SK: experimental design, collaboration in data analysis.

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Compliance with ethical standard

Conflict of interest The authors declare that they have no conflict of interest.

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